Taking soil samples for carbon measurement

Bulk density soil sampling as a step to estimate soil carbon stock, Central Kalimantan, Indonesia

Photo: World Agroforestry/Ni'matul Khasanah

Suggested citation:

Van Noordwijk M, Barrios E, Shepherd K, Bayala J, Öborn I. 2019. Soil science as part of agroforestry. In: van Noordwijk M, ed. Sustainable development through trees on farms: agroforestry in its fifth decade. Bogor, Indonesia: World Agroforestry (ICRAF) Southeast Asia Regional Program. pp 63–92.

CHAPTER FOUR

Soil science as part of agroforestry^a

Meine van Noordwijk, Edmundo Barrios, Keith Shepherd, Jules Bayala, Ingrid Öborn

Highlights

- New answers to land degradation problems have been an agroforestry focus for four decades
- Plot-level experimentation following agronomic traditions proved to be a challenge due to lateral interactions
- Testing hypotheses at process level and analysing tree-soil-crop interactions led to synthetic simulation models
- Policy attention to soil-nutrient replenishment in Africa and alternatives to slash-and-burn in humid tropics required more than technical analysis

4.1 Introduction

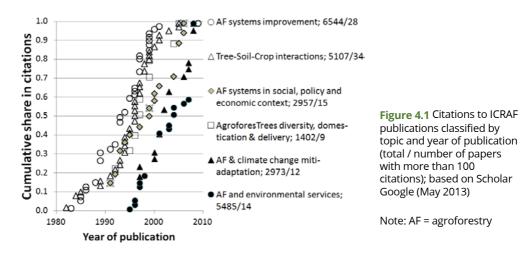
World Agroforestry (ICRAF) has as its mandate all agricultural land use that involves trees, beyond what is considered to be forest. The latter distinction is rather fluid, both temporally and institutionally, as the example of long-rotation shifting cultivation may show. Agroforestry itself ranges from croplands with a few trees added through to systems where tree crops (considered to be agricultural, such as coffee, cacao or rubber) provide a perennial vegetation layer, augmented with upper canopy layer trees utilized to modify microclimate, yielding economically valuable products. The consequences for soil conditions and functions vary along this range.

Agroforestry research has from its start operated on the active and often contested interface of the need to increase agricultural production, overall and per unit area, and the need to find more sustainable ways of managing natural resources. Agroforestry is typically associated with 'integrated' rather than 'segregated' solutions to meet the dual imperative, with specific attention to the understanding and management of trade-offs at the scales of farmers, the landscape, (sub-) national governments and the global policy arena. Soils have a key function

^a A more extensive version^a is available as ICRAF Working Paper 200

to both issues of land productivity and environmental effects, and soil research of one type or another has been part of nearly all research activities of ICRAF from its start.

Classifying the research output of ICRAF on the basis of citations to publications grouped by topic (Figure 4.1) shows six identifiable waves. Virtually all literature on agroforestry systems and improvement or 'tree-soil-crop interactions' that had been cited by 2013 had been published before 2000; by contrast, publications on agroforestry and environmental services and climate-change mitigation and adaptation started in the mid-1990s and flourished after 2000. Intermediate time patterns (steady progression in time) are found for agroforestry systems in social, policy and economic contexts, and for tree domestication.



This chapter reviews progress in agroforestry soil science in the past two decades under seven headings and provides key references for each that point to more detailed reviews and syntheses.

1. Big-issues agenda with attention to local knowledge: soil depletion, land degradation, global climate change and loss of biodiversity

One of the first documents produced when ICRAF was being formed described the issues of land degradation in the tropics and the urgency of finding solutions for intensification that combine technical, ecological, social and economic aspects. This topic remained important in the first ten years of ICRAF¹ and forms a red thread through thirty-five years of institutional history. Partial successes have not yet combined to the breakthroughs needed at global scale because the issue interacts with international terms of trade, value chains for inputs and outputs in the local, national and international economies, and the dynamic rural–urban interface and its consequences for management of food prices. Arguments for public investment in soil-fertility replenishment in Africa received attention² but they were not backed by economic policy analysis, while the technical aspects of supporting phosphorus levels so that tree and grain legumes through biological nitrogen fixation could do the job of adding nitrogen to the soil were not convincing at farmer level^{3,4}. Some success was made with fertilizer trees in fallow rotations but national subsidies for N-fertilizer to support grain crops won the day when food shortages became urgent again in southern Africa. Saving

Africa's soils still requires a combination of policy with science and technology for improved soil management that is not yet on the agenda in Africa⁵. The call for new initiatives to save Africa's soils remains urgent⁶.

The 1992 Rio conference where the primary global environmental conventions were shaped marked the start of a new interest in how local, national and international actors interacted in the process of tropical forest conversion and how changes in land-use practice could be part of a package that obtained equal local benefits but substantially reduced global impacts on climate and biodiversity⁷. The Alternatives to Slash and Burn (ASB) program was initiated to identify and support sustainable land-use intensification in the tropical forest margins, alongside protection of remaining forests. While declining soil fertility under reduced fallow lengths is one of the classical storylines that can be quantified in simple models⁸, the focus of ASB was not on traditional shifting cultivation for subsistence livelihoods⁹ but on its modern market-related versions. Almost from the start, researchers recognized that slash-and-burn as a method of land clearing was used by large-scale operators as a cheap way of establishing plantations, as a starting point for low-intensity grazing and as part of traditional shifting cultivation and crop-fallow rotations. The research program described patterns of land use in their social, economic and environmental contexts and then focused on a comparison of consequences of various land-use alternatives for an array of criteria. Soil-related constraints were found to be part of a much broader set of ecological, economic and social determinants of land-use patterns^{10,11}. This led to analysis of trade-offs and interest in the way drivers of business-as-usual change can be leveraged to nudge systems into a more desirable direction¹².

An important part of the ASB research agenda along the forest transition curve was the rehabilitation of abandoned land, as an alternative to further deforestation. There was major confusion on whether such land areas were 'degraded' or abandoned for other reasons, for example, related to tenure issues and continued forest institutional regimes that prohibited other land uses¹³. The extent and dynamics of *Imperata* grasslands in Southeast Asia were reviewed¹⁴, with specific attention to soil conditions. The latter were found to not be a real constraint to subsequent intensification^{15,16}.

Agricultural systems can greatly benefit from integrative approaches that combine formal and informal knowledge to address current sustainability problems associated with global change^{17,18}. There is increasing recognition of the potential value of knowledge held by land managers who have been closely interacting with their environment for a long time to contribute important insights about the sustainable management of natural resources¹⁹. Increased concern about soil management as a key determinant of sustainability in agricultural landscapes has driven the demand for early warning indicators to monitor changes in soil health, and their impact in the provision of ecosystem services, as affected by land-use change and agricultural intensification^{20,21}. A participatory methodology has been published recently, following several years of South–South collaboration, to guide the mobilization and integration of local and scientific knowledge on indicators of soil quality and soil fertility management²². It was designed to facilitate bottom–up approaches that integrate local knowledge into soil management decision-making processes and strengthen the relevance, credibility and legitimacy dimensions required for the adoption of best management practices. This methodological guide describes how to apply participatory tools

in identifying, classifying and prioritizing local indicators of soil health knowledge so that they can complement technical indicators and later build farming communities' consensus about how to best address soil-health constraints following agroecological management principles and integrated soil-fertility management options. The development of a 'hybrid' knowledge base, combining local and scientific knowledge, reflects an effort to understand the complexity of land management decision making to promote and protect multifunctional land uses^{23,24,25}. This is part of a continuing effort to develop land quality monitoring systems that strengthen local environmental and agricultural institutions and communities with tools that support local decision-making in natural resource management and promote sustainable land use in agricultural landscapes²⁶.



Corn-based agricultural development policy has led to land degradation in certain area in Gorontalo, Sulawesi. Photo: World Agroforestry/Ni'matul Khasanah

2. Agroforestry as a way to manage C, N, P capitals and beyond

In its first decade, ICRAF science dealt with an inventory of the diversity of agroforestry systems of the world and their primary reasons for existence. Soil and land-health management, interpreted as a combination of erosion control and maintenance of soil fertility²⁷, was identified as one of the strongest rationales for combining trees, crops and livestock on sloping lands. Soil-fertility improvement and better nutrient use-efficiency when introducing and managing trees (serving as nutrient pumps and safety nets) in agroecosystems were the focus of research aiming at optimizing agroforestry systems²⁸. From the crop's perspective, however, most trees in most circumstances have a direct negative effect based on competitive resource capture, and the longer-term benefits of inclusion of trees will only weigh up to the negatives in well-defined circumstances^{29,30}. Those circumstances potentially include, beyond sloping land, the seriously nutrient-depleted landscapes of Africa on geologically old soils^{31,32,33}.

In the 1980s, major hope became vested in alley cropping or hedgerow intercropping. Inspired by farmer-developed technology on sloping land in Flores, Indonesia, it was popularized in Africa by an Indonesian soil scientist working at the International Institute of Tropical Agriculture in Nigeria. There are many versions of the history of the hope–hype–crash dynamics of public expectations of what this technology could deliver and how lessons could be learned from the experience³⁴. It was to be largely repeated, however, in the improved fallow and fertilizer tree story that replaced it as a 'silver bullet' solution. While not ultimately leading to widespread success, the research done on hedgerow intercropping and improved fallows helped identify underlying principles on the technical, social and economic sides^{35,36,37,38,39}. The search for locally appropriate agroforestry solutions continued.

Many studies have shown that soil organic matter (SOM) content of soil under trees is higher than in soils outside tree influence⁴⁰. The attribution of this pattern to aboveground litter fall and belowground root turnover depends on local context⁴¹. However, crop yields do not correlate with total SOM, first of all due to the associated competition for light, water and nutrients but also because nutrient release from SOM is largely dependent on the fraction of SOM that is biologically active⁴². Aware of the competitive effects in simultaneous systems, research effort shifted to rotational crop-fallow systems as these are easier to understand and still part of farmers' reality. Efforts to identify biologically active fractions of SOM have shown that the amount of N in organic matter that is not physically protected and associated with soil particles, that is, light fraction N that floats on water or solutions⁴³ of densities below 1.1 g cm⁻³, can be used as a sensitive measure of differences in SOM among cropping systems⁴⁴ as it correlates with whole soil N mineralization⁴⁵. Planted tree fallows significantly modified light fraction SOM when compared to a continuous unfertilized maize control; total SOM, however, was not affected⁴⁶. Furthermore, while the amount of N in the light fraction correlated with maize yield, the quantity of light fraction SOM did not, thus, highlighting the importance of organic input quality in soil N availability⁴⁷. Key attributes of trees with the highest potential to increase soil N availability include the ability to fix nitrogen and litter with low (lignin + polyphenol)/nitrogen ratio that results in fast decomposition rates⁴⁸. Additionally, planted tree fallow studies in which SOM fractionation and sequential P fractionation were conducted on the same soil samples showed that the amount of P in the light fraction could serve as sensitive indicators of the 'readily available' soil-P pool⁴⁹. Planted tree fallows, therefore, have been successfully used to regenerate degraded soils in Africa and Latin America in areas where population pressure on land is reduced^{50,51}.

With trees as the primary point of differentiation between agroforestry and agriculture and range management, the specific aspects of perennialism imply a different sampling in space and time of soil functions^{52,53}. Trees tend to be deeper rooted (with many notable exceptions⁵⁴) and sample a much larger horizontal area, challenging traditional plot-based research despite all efforts at trenching-off plots. The net effect (positive or negative) for a farmer of inclusion of trees in an agricultural system depends on A) total resource capture (TotCapt), B) harvest index of resources captured (HarvIndex), C) losses to other environmental compartments of resources not harvested (Loss), D) economic value of the resources harvested (Value) and costs of losses to the environment (Cost), E) the expenditure of labour and other inputs at going price (Price) and F) possible changes in land value (ΔLandValue):

NetBenefit = TotCapt * HarvIndex * Value – Loss * Cost – Inputs * Price + ΔLandValue

Research has tried to dissect this by relating A to tree architecture, phenology and growth rate, potentially independent of B and D, which are the focus of tree domestication and tree improvement efforts, alongside value-chain economics. Aspect E, the labour requirements of keeping the competitive aspects of trees under control while benefitting from the positive contributions to local nutrient cycles, proved to be a major challenge for the once-popular hedgerow intercropping systems. Meanwhile, aspect C has gained importance with current refocus on greenhouse-gas emissions, alongside erosion and leaching losses of soil particles and solutes. Aspect F may still be under-researched.

After a period of intensive research at process level on total resource capture, the conditions where 'over-yielding' of mixtures involving trees are fairly well established, while the effects of trees on losses by erosion and greenhouse-gas emissions have been quantified for a range of situations⁵⁵. The interactions between trees and soil biota have been well explored in terms of mycorrhiza and earthworms (as reviewed later in this chapter) but a large part of the soil biological spectrum is open for further discovery. Science-based perspectives on bio-economic modelling can be compared with farmers' preferences and knowledge in the joint design of new management systems.

With depletion of agricultural soils due to nutrient export beyond the replenishment by fertilizer identified as a key challenge of farming⁵⁶, especially in Africa^{57,58}, considerable effort has been directed to the use of trees as 1) sources of biologically fixed nitrogen^{59,60}, 2) recyclers and safety-nets of nutrients from deeper layers⁶¹, and 3) converters of less-processed nutrient sources, such as rock phosphate. However, farm-level nutrient budgets^{62,63} caution that agroforestry can result in large nutrient extractions in product removals while pointing to opportunities for nutrient imports through livestock feed. The potential for tree fallows to re-capture leached nitrate held in the subsoil on anion exchange surfaces was demonstrated^{64,65,66} and also the ability to reallocate some of the soil P into more labile P-pools⁶⁷. While a number of technical solutions have emerged that are still worth further testing⁶⁸, no silver bullets have emerged that could revolutionize farming under the constraints of high nutrient exports and low economic feasibility of input use. As an alternative direction, the shift to tree crops with high economic value per unit harvested product has proven to be more successful.

Complementing the process and modelling approaches, new efforts are currently being made to efficiently describe the spatial variation in soil properties in the hope that this can lead to better targeting of sustainable land management practices while allowing for monitoring at real scale how soil properties change in response to land use⁶⁹. A major challenge for any quantification of 'impact' is the counterfactual: what conditions could be expected without the intervention that is evaluated for impact? Any comparison of current soil conditions under two land-use systems must account for possible *a priori* differences between the locations where the two systems developed. This requires understanding of the existing variation in the landscape, local knowledge of conditions, preferences for specific parts of the landscape for specific land uses and ability to implement preferences⁷⁰. There are some examples of tightly controlled designs for assessing changes in soil conditions in landscapes for forest transitions ⁷¹ and exclosures⁷².

3. From process hypotheses and plot-level experiments to synthetic tree-soil-crop interaction models and management of filter functions

Research on tree-soil-crop interactions in agroforestry have focused on growth resources sharing between trees and crops mediated by soil with the hypothesis of trees creating favourable microclimatic and soil modifications for the crops. The findings have shown that trees on farms provide services to agriculture by contributing to 1) extended growing season by keeping the landscape covered with vegetation, 2) regulating water flows to the benefit of crops and groundwater recharge, and 3) soil regeneration, carbon sequestration and nutrient cycling⁷³. However, the potential benefits depend on complex spatial and temporal interactions between the biological, physical, hydrological and climatic components of the system^{74,75}. Such interactions change with time as trees grow larger together with the processes that affect the soil, which are governed by the root systems to a large extent^{76,77}, and also by the tree phenology^{78,79,80}. Finally, management practices also affect these interactions, such as the tree density and vegetation cover, the use of fires to clear land^{81,82,83}, pruning of tree crowns or roots^{84,85,86}, and the maintenance of pruned biomass, crop residues and litter as mulch^{87,88,89,90,91}.

While tree species vary in rooting architecture and root biomass, tree roots can extend to deeper soil layers compared to those occupied by crop roots. They may, therefore, take water from the groundwater even though there is evidence of trees taking water from the top soil layers as well, depending on the species and its root systems⁹². Nevertheless, it is worthwhile to mention that there is no direct relationship between tree water extraction and fine root density because decreasing water potential also plays a role⁹³. The effects of the increase of CO₂ and temperature as a result of climate change on changes to soil carbon storage were reported to be contradictory, calling for more investigation to separate the effect of increased C and that of possible changes in roots and rhizospheres⁹⁴. In mixed agroforestry systems, the use of isotopes has helped to disentangle the contributions of the components and revealed larger contribution of the C3 plants (trees) to soil carbon in comparison with annuals^{95,96,97}. As tree roots can grow deeply, they can also lift water and, with it, nutrients that leached below the reach of crops. They can act as a 'safety net' to capture nutrients leached from the topsoil and redistribute them to the soil surface^{98,99}. Such a mechanism has been reported to improve N use-efficiency¹⁰⁰. In addition, estimates of water volume lifted/redistributed can represent up to 30% of daily evapotranspiration^{101,102}. According to the authors, this has a number of eco-physiological implications, for example, maintaining fine root viability and resistance to drought while affecting some of the soil processes, such as increasing soil water and soil biota activity.

Some synthetic analyses of published data using meta-analysis have also helped understand in which circumstances soil improvement translates into better crop production^{103,104}. Another review and meta-analysis¹⁰⁵ showed that spatial heterogeneity in savannah vegetation was a result of termite mounds being fertility spots in the landscape, enriched with clay, carbon, nitrogen, calcium, magnesium and potassium.

Field investigations have helped generate a wealth of information on processes in isolation but have failed to reveal which one was the most prominent. A solution to this problem has been the development of a modelling phase which tried to synthesize the generated information to reveal the most limiting factors and processes for the production of associated crops. For instance, simulations using Water, Nutrient and Light Capture in Agroforestry Systems (WaNuLCAS)¹⁰⁶ revealed that the decrease in *Zea mays* growth near *Grevillea robusta* water was due to lower soil-water content that resulted in a decreased P diffusion¹⁰⁷.

Similarly, water was found to be the most limiting nutrient under *Vitellaria paradoxa* while it was P under *Parkia biglobosa¹⁰⁸*. For planning adaptation, WaNuLCAS was also used¹⁰⁹ to evaluate the effects of different management options (tree density, tree pruning, mulching and root pruning) on *Sorghum bicolor* production under future climate scenarios. There are a number of other models (APSIM, HiSafe, HyPAR, SCUAF etc) but they all have their limitations that are inherent to models, such as over-simplification, or our poor understanding of the processes involved in tree-soil-crop interactions, or to both¹¹⁰. If combining field investigations and modelling has helped to generate some scientific advances, there are still some methodological challenges in determining the 'parkland effect' (effect of a group of trees on biodiversity, microclimate etc), the trade-offs and synergies between and among goods and services, and how to boost the provisioning of ecosystems services¹¹¹.

Empirical research on agroforestry was initially largely built on the agronomic traditions of replicated field trials with plots in which a border zone was excluded from yield measurements to minimize lateral interactions between plots. Root research on trees, however, revealed that for many trees the lateral expansion can be multiples of the canopy height¹¹² and much of the experimental evidence needed to be interpreted with caution. It is possible that 'control' plots were effectively mined by tree roots from neighbouring plots, the performance on such plots enhanced by external nutrient capture, and hence the contrast between plots with and without trees magnified. Digging (deep) trenches around plots brings only temporary relief, as tree root systems can within a year occupy the space. A welldesigned, replicated field trial¹¹³ on various types of hedgerows as erosion control strategies showed that the underlying variability of the hill slope with respect to infiltration capacity had a major effect on what was measured as overland flow at plot level and the effectiveness of hedgerows as filters depended on the position of measurement. Much of the subsequent research relied on understanding spatial variability in the field rather than on controlled experiments. Despite substantial effort to spatially parameterize the Universal Soil Loss Equation (USLE) for a landscape in Kenya¹¹⁴, the model correctly classified only 38% of sites into three degradation classes and the model sensitivity for delineating regions of severe degradation was only 28%. Local calibration with ground data could increase the correctly classified sites to 54% but without expectation that a modified model would be valid elsewhere. Similarly, there was little spatial agreement between predictions of different models (including modified USLE approaches) for a coffee-dominated landscape in Lampung, Indonesia¹¹⁵; there, in-field erosion was found not to be the major determinant of river sediment transport. Overland sediment flows were partially filtered while paths used for motorcycles, roads and shallow landslides contributed sediment directly to the river. Sediment and soil transport issues appeared to have different determinants at every scale between a soil pedon, a plot, a hill slope, a small and a large catchment. The fractal dimension that characterizes net sediment transport with a length scale to the power 1.5–1.6 was found^{116,117} to have a parallel in the social organization of watershed management institutions¹¹⁸. There has been little accompanying work on the economic costs of soil erosion

and benefits of agroforestry. Estimated ecological–economic costs of soil erosion in Kenya using emergy analysis at the national level were found to be equivalent to the value of agricultural exports or electricity production¹¹⁹.

A further step in the scientific understanding of agroforestry came when lateral resource capture was seen not only as a challenge to research aimed at defining technology for 'homogenous' conditions but as an important aspect of real-world agroforestry, especially in the mixed stands typical of smallholdings, where edge planting of 'aggressive' trees may imply that half of the nutrients are scavenged off farm. This perspective on lateral resource capture aligned with the analysis of hedgerows of trees and naturally vegetated strips on sloping land. Rather than defining a uniform technology, science helped articulate a perspective on a range of niches in a diverse farming environment, with variation in tree properties that can be understood in, and used in fine-tuning of, farmers' decisions to plant, prune, manage, harvest and/or remove¹²⁰.

4. Trees and other soil biota: old tree root channels, earthworms, mycorrhiza, rhizobia and nematodes

Trees live above as well as belowground. Soil structure is a key determinant of root development and root function, as well as for other soil biota. Soil compaction as a consequence of agricultural use and/or overgrazing is both a symptom of soil (mis-)management as well as a cause of declining primary productivity. The importance of this, however, varies with the rainfall regime and climate zone. Macroporosity of soils, the class of pores most easily compacted, is essential for saturated hydraulic conductivity and the ability of soils to handle intense rain without overland flow and ensuing erosion. Macroporosity in the field is linked to texture (cracking clay soils), decayed tree root channels¹²¹, the impact of deeply burrowing earthworms¹²² and possibly other soil biota. Measurement of infiltration in the field typically shows log-normal distributions, with a small fraction of points having one or two orders of magnitude of higher infiltration rates. The question of how such infiltration hotspots at field scale operate during extreme rain events cannot be easily assessed from current measurement techniques because much depends on their subsoil connectivity to landscapelevel drainage systems. Agroforestry can influence the continuous formation of macroporosity through the provision of leaf-litter-feeding earthworms and, at another time scale, the formation of decaying tree root channels. At the level of mesoporosity, the tendency of soils to form aggregates is strongly influenced by soil-ingesting soil biota¹²³ and by fungal hyphae associated with mycorrhiza¹²⁴. Attribution of biological activity associated with soil structure modification is not a trivial exercise but a methodological approach using Near Infrared Spectrometry (NIR) allowed the separation of soil aggregates produced by soil invertebrates and roots living in the same soil¹²⁵.

Vertical and horizontal water transport through and over the surface of soils is, however, a 'communicating vessels' problem with strong trade-offs. If water flows over the surface it may cause erosion but it reduces the problem of leaching, and vice versa. A more detailed examination of by-pass flow, however, made clear that macroporosity can drain excess water without much effect on solute transport in mesopores, especially where the latter benefit from physico-chemical ion adsorption acting as an additional safety net¹²⁶. Later versions of

the WaNuLCAS model¹²⁷ have included such processes and allow the dynamics of soil structure, bypass flow and root-based safety nets for leaching nutrients to be quantified.

In Burkina Faso, with yearly rainfall ranging 570–1180 mm, groundwater recharge was simulated to be the equivalent of 2–14% of the total gross water input. A combination of the measurement and modelling of drainage and transpiration in agroforestry parkland revealed that intermediate density of trees (5–25 trees ha⁻¹ based on the assumption that 100–0% of transpired water came from below 1.5 m depth) can maximize groundwater recharge while at higher stockings there was a trade-off between tree cover and available water¹²⁸.

The soil environment may well host and interact with the most complex biological community once we account for scale¹²⁹. Soil biota (for example, microbes, invertebrates) mostly contained in the upper few decimetres of soil are extremely diverse and make important contributions to a wide range of ecosystem services that are essential to the sustainable function of natural and managed ecosystems^{130,131}. New high-throughput DNA profiling techniques are supporting efforts to assess the global distribution of soil biota and the relationship of belowground biodiversity to above-ground biodiversity¹³². Soil biota directly influence soil fertility by mobilizing nutrients¹³³ and form soil structures¹³⁴, increasing water infiltration and soil C storage and decreasing soil erosion. Therefore, in order to understand the distribution and diversity of soil organisms and how they respond to disturbance, be it agricultural practices or climate change, it is necessary to monitor the soil and environmental quality that is required for sustaining land health in agricultural ecosystems¹³⁵. Strategies for maintaining native biota of farm soils, such as mycorrhizal inoculum potential, are generally preferable to inoculation strategies^{136,137}. Recent global studies show that preservation of plant biodiversity is crucial to maintain multiple ecosystem functions like nutrient cycling, plant productivity and carbon storage, and also to buffer the negative effects of climate change¹³⁸. Slash-and-mulch agroforestry systems show greater abundance of soil macrofauna than native forest, suggesting that maintenance of soil cover with organic materials of different qualities promotes favourable conditions for soil biological activity¹³⁹. Comparison of adjacent agricultural plots with and without trees show that tree presence increases abundance of several groups of soil biota¹⁴⁰. Further, greater soil biological activity occurs near trees but the effect is greater for some tree species than for others¹⁴¹. This is likely related to differences in plant functional traits¹⁴². Trees can be considered as 'hot spots' of biological activity and play a major role in maintaining and promoting soil biological activity responsible for many of the functions that underpin soil-mediated ecosystem services¹⁴³. Farmers' perspectives and knowledge of soil biota together with scientific knowledge contributes to a better understanding of tree-soil biota interactions in time and space that allow design of diverse cropping systems that can sustain multiple functions required for the adequate provision of ecosystem services^{144,145,146,147}.

5. Soil-carbon dynamics and greenhouse-gas emissions from agroforestry systems

The ASB program was the first to establish a cross-continental network of sites with consistent measurement of above- and belowground carbon stocks of forests and forest-derived land uses in the humid tropics¹⁴⁸. A review of the way soil carbon stock varies with soil type, elevation (temperature) and land cover introduced the concept of C-reference values and associated soil carbon deficits¹⁴⁹, taking natural forest soils with the same texture, mineralogy,

pH and elevation as the basis for a pedotransfer function. The empirical relationships between texture and soil-carbon content were aligned with the assumptions and process descriptions of the Century model; attempts to measure the 'functional' fractions represented in the model remained partially unsuccessful, however¹⁵⁰. Analysis of carbon dynamics in aggregate fractions¹⁵¹ could not be directly linked to fully functional carbon-balance models.



Farmers in Na Thau village, Veitnam, take samples in their community forest for soil-carbon measurement. Photo: World Agroforestry/Duc Minh Nguyen

Carbon stocks are additive and allow area-based scaling, making it straightforward to scale from plot to landscape¹⁵² once the scale-dependent patterns of spatial variation are known. The high spatial variability of soil carbon, coupled to costs of sampling and analysis and challenges in attributing differences to cause–effect chains, make it unlikely that soil carbon, when assessed with current standard methods, can become part of carbon projects¹⁵³. More optimistic perspectives related to methodological improvements will be discussed below. A further challenge to such inclusion is the observation that a 'soil-carbon transition curve', with recovery following degradation, can be observed in response to agricultural intensification, and without specific soil-carbon interventions¹⁵⁴. Rather than being a primary target for interventions and finance as part of climate-change mitigation, soil carbon should be of interest from the perspective of buffering of soil water and nutrient content, as part of farm resilience and climate-change adaptation¹⁵⁵.

The early measurements of nitrous oxide and methane emissions in relation to tropical landuse change suggested that such fluxes will generally be small relative to the greenhouse-gas effect of tropical forest conversion through changes in (mostly aboveground) carbon stocks. Specific for the use of N₂ fixing shrubs and trees in agroforestry, where N-rich mulch is left on the soil surface without incorporation into the soil, high emissions of nitrous oxide are possible and were measured in shaded coffee systems¹⁵⁶. In terms of net greenhouse-gas effects, the jury is out to determine whether biological N_2 fixation by trees is friend or foe¹⁵⁷; the likely answer is that it depends on how and where such trees are used.

6. Soil/land health surveillance

ICRAF's work on low-cost, rapid, soil characterization using diffuse reflectance spectroscopy began with the use of field spectroscopy in combination with Landsat imagery to trace sources of soil erosion in Lake Victoria¹⁵⁸. This early work, using the visible-near-infrared (VNIR) wavelength range, also showed the potential for using soil reflectance to measure management-induced changes in soil quality in long-term trials¹⁵⁹. This was later demonstrated at landscape scale in land-use change studies in Madagascar¹⁶⁰ and along a tropical forest-cropland chrono-sequence in Kenya¹⁶¹. A scheme for the use of spectral libraries as a tool for building risk-based approaches to soil evaluation was demonstrated for a diverse library of over 1000 topsoils from eastern and southern Africa, including development of spectral diagnostic tests for screening soils with respect to critical soil-fertility limits¹⁶². The global applicability of soil spectroscopy was further demonstrated using a global soil library based on archived samples at the US National Soil Survey Center using VNIR¹⁶³ and for available samples from the International Soil Reference and Information Centre global archives using mid-infrared spectroscopy¹⁶⁴.

Infrared spectroscopy uses a different set of principles than conventional soil-fertility tests, providing a single multiple-utility measure of soil-production potential and response to management^{165,166}. With IR, soils can be characterised in a single 30-second measure that requires no chemicals, only light. The shapes of infrared spectra respond to the basic molecular structure of mineral and organic composition of soils and their interactions. It is the organic–mineral composition that determines soil functional properties, including a soil's ability to retain and supply different nutrients and water, nitrogen mineralisation capacity, soil charge characteristics, soil structural stability and ability to resist soil erosion, and amount of soil organic carbon in different pools and its protection. Although calibration to conventional soil tests has been used as an intermediate step, the ultimate concept behind the spectral approach is to calibrate soil and crop responses to management directly to infrared spectra and completely by-pass the need for conventional soil tests¹⁶⁷.

The ability to derive spectral indicators of soil fertility was demonstrated in several studies¹⁶⁸ that successfully calibrated soil-condition classes, based on ten commonly used agronomic indicators of soil fertility, to both soil reflectance measured in the laboratory and Landsat TM reflectance, which permitted mapping of the index. The spectral index also related to δ 13C dynamics associated with historic land-use changes, similar to other studies that spectrally discriminated forest–cropland chronosequence classes¹⁶⁹. A similar approach was successfully used for spectral prediction and mapping of soil-fertility classes in Mali¹⁷⁰, while a similar study¹⁷¹ calibrated principal components of soil-fertility variables to spectra to assess the prevalence of soil-fertility constraints on farm fields in Kenya.

Several studies have shown strong relationships between observed or measured soil erosion in the field and laboratory-measured soil spectra. A study in Kenya's Nyando River Basin¹⁷² was able to spectrally discriminate ground visual observations of three ordinal erosion classes with validation accuracies of 78%. A similar approach in Kenya's Saiwa River Basin¹⁷³ obtained

validation accuracies of 72%, with additional validation of the erosion classes using soil ¹³⁷Cs concentration data. An erosion-deposition index was developed ¹⁷⁴ as a tool to rapidly screen soils in the Nyando River Basin into eroded, intact or depositional soil classes based on a spectral distance index using sediment spectra as a reference library. The spectral index was validated using ¹³⁷Cs analysis and soil spectra were also used to interpolate ²¹⁰Pb concentration in sediment cores. The combined data allowed a sediment budget for the basin to be constructed as well as the historic time trends in soil erosion from 1900.

Soil spectroscopy was shown to be able to predict various soil-carbon fractions and their mineralization rates. Mid-infrared (MIR) spectroscopy was used to predict the concentration of organic-carbon fractions present in a diverse set of Australian and Kenyan soils¹⁷⁵. The coefficient of determination of measured versus predicted data (r^2) ranged from 0.97 and 0.73 for total organic carbon, particulate organic carbon, and charcoal carbon. Soil spectra were also shown to predict carbon mineralization rates from different soil physical fractions in two contrasting soil types¹⁷⁶. At the same sites, mid-infrared spectra were used to interpret functional groups to help elucidate biogeochemical mechanisms that determine the fate of carbon inputs in soils and organic matter stabilization by aggregates¹⁷⁷. Removing the mineral soil spectra in Alfisols, obtained from heated soils, did not improve spectral calibrations of soil organic carbon, indicting the robustness of the spectral method¹⁷⁸.

Reflectance spectroscopy was shown to be useful for predicting organic resource quality for soil and livestock management based on nitrogen, lignin and soluble polyphenol concentrations^{179,180}. Validation r² of >0.8 were obtained for prediction of in vitro dry matter digestibility (IVDMD) and C and N mineralization for a diverse range of crop and tree residues of varying quality^{181,182,183}. NIR for determination of crude protein content in cowpea (*Vigna unguiculata*) leaves was also demonstrated¹⁸⁴.

Infrared spectroscopy can enable an evidence-based diagnostic surveillance approach to agricultural and environmental management in developing countries¹⁸⁵, based on the scientific principles used in public health surveillance. Infrared spectroscopy was proposed as a rapid screening tool for assigning samples to case or reference and allowing characterization of the health of systems at scale using population-based sampling. The diverse range of applications of infrared spectroscopy in agriculture and environment was reviewed.

In response to the need for objective, quantitative and cost-efficient methods for assessment of land health to justify, target and prioritize investments, the diagnostic surveillance principles were taken further to form a conceptual framework for wide-area soil and landhealth surveillance^{186,187}. Land health is defined as the capacity of land to sustain delivery of ecosystem services and is a prerequisite for wise ecosystem management and sustainable development. The soil spectroscopy methods were key to enabling this approach by providing a soil analytical tool that could be applied cost-effectively at scale. Land-health surveillance is hinged on systematic georeferenced field observations based on probability sampling (Land Degradation Surveillance Framework/LDSF)^{188,189}, so that inferences can be made back to the target area sampled. Georeferenced soil spectral estimates of soil properties are statistically modelled to remote sensing covariates so that the models can be applied back to every pixel on the satellite imagery to provide digital maps of soil properties. The report and accompanying atlas¹⁹⁰ illustrate the land-health surveillance concepts with a case study in the West African Sahel, presenting results on regional remote-sensing studies of historical changes in vegetation growth and rainfall patterns in the area, indicating land-degradation trends, and on field-level assessment of land degradation in Mali. This combination of principles and scientific and technical advances formed the basis for the Africa Soil Information Service (AfSIS).

ICRAF played a foundational role in the establishing AfSIS. The project has implemented the first ever probability sample of African land health and soils, based on a set of 60 100-km² sentinel sites, providing a baseline for future monitoring of soil-health changes (<u>www.africasoils.net</u>)¹⁹¹. Spectral measurements were performed on all samples, while conventional reference measurements were done on a 10% random subsample¹⁹². A centralized African soil spectral prediction service is being piloted based on Bayesian Additive Regression Trees. This will allow spectrometer users to submit batches of spectra online and obtain predictions of soil properties with uncertainties given for each sample. Samples that are spectral outliers or have large prediction error can be submitted to the ICRAF laboratory for characterization and adding to the calibration library. This service could drastically reduce the need for conventional soil testing.

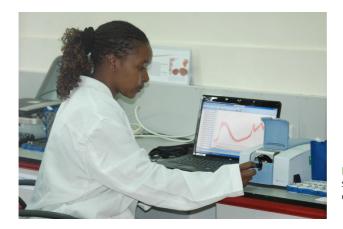


Figure 4.2 Portable mid-infrared spectrometer being used for rapid characterization of soil samples

In support of this initiative, ICRAF established a globally unique, Soil-Plant Spectral Diagnostics Laboratory, which focuses on analysing soils using only light (infrared, x-ray, laser). The laboratory established Fourier Transform near- and mid-infrared spectroscopy as a foundation for calibration transfer across a network of spectrometers. The light-based technologies have been extended to: benchtop x-ray diffraction for mineralogical analysis; total x-ray fluorescence for total element analysis¹⁹³ in soils, plants and water; handheld x-ray fluorescence spectroscopy; and laser diffraction particle size analysis for dry and wet aggregate stability, for which standard operating procedures are available at: <u>http://worldagroforestry.org/research/land-health/spectral-diagnostics-laboratory</u>. The laboratory supports a soil spectroscopy network spanning 10 African countries, to which it provides scientific and technical backstopping, including on-site training. Extensive support has been provided towards the establishment of the Ethiopia Soil Information System (<u>http://www.ata.gov.et/projects/ethiopian-soil-information-system-ethiosis/</u>). To enable easier access to soil spectral calibration techniques, ICRAF has developed the soil.spec software package in R (<u>http://cran.r-project.org/web/ packages/soil.spec/soil.spec.pdf</u>) and now runs an international soil spectroscopy training course.

Land-health surveillance approaches supported by soil spectroscopy are being applied in a number of sustainable land management projects in 10 African countries and in the CGIAR pan-tropical sentinel sites initiative. These include applications such as mapping soil carbon in rangelands¹⁹⁴, monitoring and degradation prevalence and soil functional properties in Ethiopia¹⁹⁵, and studying patterns in soil faunal and microbial activity in landscapes¹⁹⁶. Soil spectroscopy has also been used to characterize patterns of variability in soil fertility in smallholders' farming systems^{197,198,199,200}. Current applications include a pilot on integrating monitoring of soil fertility on farms into the World Bank Living Standards Measurement Study and soil monitoring in an integrated monitoring system for ecosystem services in agricultural landscapes (<u>www.vitalsigns.org</u>). Soil spectroscopy is also now being used by two private soil-testing services in Kenya.

Systematic application of land-health surveillance has the potential to generate improved understanding and predictive ability of agricultural systems and natural resources at multiple scales and improve intervention decision planning and impact assessment. Technological advances will lead to reliable handheld and mobile phone-based spectrometers and put the technology in the hands of farmers. The CGIAR can play an important role in building up centralized, online spectral calibration and advisory services. Digital mapping techniques based on Bayesian spectral-spatial one-step modelling with prediction uncertainties generated are already in development. These scientific and technical advances are paving the way for a new paradigm of predictive agronomy and crop breeding whereby response trials are co-located with soil-spectral measurements and remote-sensing observations. This could greatly enhance our ability to predict and map uncertainty in responses to soil and crop management and perhaps by-pass conventional soil tests. While the biophysical understanding of soil management has received much attention, there is need for much more attention on demonstrating the economic value of soil ecosystem services and improved soil management practices, and to better integrate soil information into decision-making processes^{201,202}.

7. The challenge of demonstrating development impact through soil changes

While the balance that draws us towards direct solutions for urgent problems of poverty, food security and environmental destruction swings back periodically to the equally pressing needs of scientific rigor and generalizable public goods, ICRAF as a CGIAR research centre has a long history of trying to satisfy all and debating where the best position is along the curve. Rather than choosing one point, it is important that the balance can swing.

From a time when 'packaged technology' was seen as a generic answer to local development challenges of many farmers in many places, we have moved forward to a greater appreciation of diversity. Spatial variability and diversity have often been seen as a problem in that they do not allow simplistic perspectives on scaling up. As 'homogeneity' has often been used as a site-selection criterion for field experiments, because it increases the chance of 'statistically significant' treatment effects to be seen with practically feasible levels of replication, scientists reviewing experimental evidence have a biased view of the world²⁰³. Technologies that were

carefully packaged by scientists are generally unpacked by farmers, who will adopt the parts they like and find new ways around the parts they don't^{204,205,206,207}. Having learnt from this experience, science and extension bodies developed a more modest approach to presenting a basket of options, with attention to risk management and the question of how many eggs should be put into each basket²⁰⁸.

Unfortunately, the funders of international agricultural research are fascinated by the numbers of farmers and the area of land that can be claimed to benefit from 'improved practices' and are linking funding decisions to a 'beauty contest' among alternative programs judged on claims to impact. A direction that offers that one 'can eat development cake' and have good science as well, is seen to be in 'research in development'²⁰⁹, with a focus on fine-tuning the baskets of options to what might have a chance to be accepted, and an equal attention to what and how farmers choose and why they do so, with social and gender stratification replacing the abstract, 'standard farmer' perceived in the past. This gives an even greater weight to taking local knowledge seriously: not only does it point to empirical experience from which formal science can learn; it also suggests a language in which scientific findings can be communicated back, alongside the baskets of options. Science in that perspective can be useful by testing and validating simple decision trees at component level²¹⁰.

 Diagnostics:
 description and quantification of current state and variation in key characteristics

 Variation in environmental, social and economic conditions that require adjustments and adaptation as part of a 'scaling out' of solutions that worked locally

 Theories of Change:
 using currently valid understanding of relationships to guide sustainable development, while remaining alert to possible discrepancies and 'red herrings'

 Changes through positive and negative feedback of system properties in response to different scales of application in 'scaling up'

 Change of Theories:
 updating understanding of relationships that guide development on the basis of new insights and findings

Figure 4.3 Key concepts of research in development, which require continued diagnostics as part of monitoring and evaluation and sentinel approaches, explicit theories of change that address both variation in circumstances encountered in 'scaling out' and changes in dynamic properties as a response of 'scaling up', and that may lead to change of theory

Change in soil properties tends to be slow compared to aboveground changes, and this 'slow variable' characteristic has consequences for impact studies. On one hand, it implies that changes in soil conditions, whether negative (depletion, degradation) or positive (restoration), once set in motion can be expected to have long lasting, negative or positive, effects that add to the importance of observed trends. On the other hand, slow change, combined with the high inherent spatial variability of soils, makes it difficult to obtain convincing evidence of any change at all. A simple spreadsheet model²¹¹ (Figure 4.4) illustrates how a sampling of soil conditions found under different land-use systems can lead to strongly biased conclusions about 'effects of land use on the soil' if it does not account for the degree to which local variation of soil conditions informed land-use patterns in the first place.

Positive or negative changes in soil conditions in response to business-as-usual development, modified by specific development interventions, impact on many stakeholders. The most

obvious ones are literally downstream, as the soil controls the switch between overland flow, with associated flashiness of rivers, and infiltration for slower 'interflow' in saturated soils and groundwater replenishment in other situations. The contrasting interests between 'water harvesting', where overland flow is to be stimulated and used, versus beneficiaries of infiltration has been noted before. The recent discourse on 'rainbow water' suggests that there are also 'downwind' stakeholders, whose interests may differ from those downstream²¹².

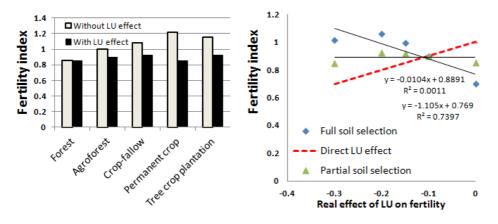


Figure 4.4 Illustration of the way land use (LU) effects soil properties, together with the preferential positioning of land uses in specific parts of a landscape, and influence survey results, with the possibility of apparent effects having opposite signs to real ones²¹³

Further progress in soil science at the ICRAF will have to address the multiple agendas of global articulation of the ambitions for sustainable development, with growing evidence that forms of agroforestry can support many of the goals²¹⁴, national green economy ambitions with land uses that minimize damage or restore soils after phases of degradation²¹⁵, and farmer's preferences and choices. The complex involvement of multiple actors in what is perceived to be 'sustainable'²¹⁶ suggests that a close linking of technical and social expertise will remain important for impact-oriented fundamental soil science in agroforestry.

References

- ¹ Sanchez PA. 1987. Soil productivity and sustainability in agroforestry systems. In: Steppler HA, Nair PKR, eds. *Agroforestry: a decade of development*. Nairobi, Kenya: International Council for Research in Agroforestry. pp 205–223.
- ² Sanchez PA, Shepherd KD, Soule MJ, Place FM, Buresh RJ, Izac AMN, Woomer PL. 1997. Soil fertility replenishment in Africa: an investment in natural resource capital. Replenishing soil fertility in Africa. pp 1–46.
- ³ Soule MJ, Shepherd KD. 2000. A regional economic analysis of phosphorus replenishment for Vihiga Division, western Kenya. *Agricultural Systems* 64:83–98.
- ⁴ Shepherd KD, Ohlsson E, Okalebo JR, Ndufa JK. 1996a. Potential impact of agroforestry on soil nutrient balances at the farm scale in the East African Highlands. *Fertilizer Research* 44:87–99.
- ⁵ Shepherd KD. 2007. Saving Africa's soils: science and technology for improved soil management in Africa. Nairobi, Kenya: World Agroforestry Centre (ICRAF).

- ⁶ Swift MJ, Shepherd KD. 2007. Saving Africa's soils: science and technology for improved soil management in Africa. Joint NEPAD, ICRAF, TSBF-CIAT Publication. Nairobi, Kenya: World Agroforestry Centre (ICRAF).
- ⁷ Sanchez PA. 1995. Science in agroforestry. *Agroforestry systems* 30:5–55.
- ⁸ Van Noordwijk M. 2002. Scaling trade-offs between crop productivity, carbon stocks and biodiversity in shifting cultivation landscape mosaics: the FALLOW model. *Ecological Modelling* 149:113–126.
- ⁹ Van Noordwijk M, Mulyoutami E, Sakuntaladewi N, Agus F. 2008. Swiddens in transition: shifted perceptions on shifting cultivators in Indonesia. Bogor, Indonesia: World Agroforestry Centre (ICRAF) Southeast Asia Regional Program.
- ¹⁰ Van Noordwijk M, Murdiyarso D, Hairiah K, Wasrin UR, Rachman A, Tomich TP. 1998. Forest soils under alternatives to slash-and-burn agriculture in Sumatra, Indonesia. In: *Soils of tropical forest ecosystems.* Heidelberg, Germany: Springer. pp 175–185.
- ¹¹ Van Noordwijk M, Hairiah K, Woomer P, Murdiyarso DM. 1998. Criteria and indicators of forest soils used for slash-and-burn agriculture and alternative land uses in Indonesia. The contribution of soil science to the development of and implementation of criteria and indicators of sustainable forest management. pp 137–154.
- ¹² Murdiyarso D, van Noordwijk M, Wasrin UR, Tomich TP, Gillison AN. 2002. Environmental benefits and sustainable land-use options in the Jambi transect, Sumatra, Indonesia. *Journal of Vegetation Science* 13:429–438.
- ¹³ Minang PA, van Noordwijk M, Kahurani E, eds. 2014. *Partnership in the tropical forest margins: a 20-year journey in search of alternatives to slash-and-burn*. Nairobi, Kenya: World Agroforestry Centre (ICRAF).
- ¹⁴ Garrity DP, Soekardi M, van Noordwijk M, de la Cruz R, Pathak PS, Gunasena HPM, van So N, Huijun G, Majid NM. 1997. The Imperata grasslands of tropical Asia: area, distribution and typology. Agroforestry Systems 36:3–29.
- ¹⁵ Santoso D, Adiningsih S, Mutert E, Fairhurst T, van Noordwijk M. 1996. Soil fertility management for reclamation of *Imperata* grasslands by smallholder agroforestry. *Agroforestry Systems* 36:181–202.
- ¹⁶ Van Noordwijk M, Hairiah K, Partoharjono S, Labios RV, Garrity DP. 1996. Food-crop-based production systems as sustainable alternatives for Imperata grasslands? *Agroforestry Systems* 36:55–82.
- ¹⁷ Joshi L, Shrestha PK, Moss C, Sinclair FL, Noordwijk MV, Cadisch G & Ong CK. 2004a. Locally derived knowledge of soil fertility and its emerging role in integrated natural resource management. In: van Noordwijk M, Cadisch G, Ong CK, eds. *Belowground interactions in tropical agroecosystems*. Wallingford UK: CAB International. pp 17–39.
- ¹⁸ Joshi L, Schalenbourg W, Johansson L, Khasanah N, Stefanus E, Fagerström MH, van Noordwijk M. 2004. Soil and water movement: combining local ecological knowledge with that of modellers when scaling up from plot to landscape level. In: van Noordwijk M, Cadisch G, Ong CK, eds. *Belowground interactions in tropical agroecosystems*. Wallingford UK: CAB International. pp 349–364.
- ¹⁹ Barrios E, Trejo MT. 2003. Implications of local soil knowledge for integrated soil management in Latin America. *Geoderma* 111:217–231.
- ²⁰ Barrios E, Delve RJ, Bekunda M, Mowo J, Agunda J, Ramisch J, Trejo MT, Thomas RJ. 2006. Indicators of soil quality: A South–South development of a methodological guide for linking local and technical knowledge. *Geoderma* 135:248–259.
- ²¹ Sileshi G, Akinnifesi FK, Ajayi OC, Chakeredza S, Kaonga M, Matakala PW. 2007. Contributions of agroforestry to ecosystem services in the Miombo eco-region of eastern and southern Africa. *African Journal of Environmental Science and Technology* 1:68–80.
- ²² Barrios E, Coutinho HL, Medeiros CA. 2012a. *InPaC-S: Participatory Knowledge Integration on Indicators of Soil Quality. Methodological Guide*. Nairobi, Kenya: World Agroforestry Centre (ICRAF); Brasilia, Brazil: Embrapa; Cali, Colombia: Centro Internacional de Agricultura Tropical.
- ²³ Sinclair FL, Joshi L. 2000. Taking local knowledge about trees seriously. Forestry forest users and research: new ways of learning. Wageningen, The Netherlands: European Tropical Forest Research Network. pp 45–58.
- ²⁴ Sileshi GW, Nyeko P, Nkunika PO, Sekematte BM, Akinnifesi FK, Ajayi OC. 2009. Integrating ethnoecological and scientific knowledge of termites for sustainable termite management and human welfare in Africa. *Ecology and Society* 14:48.

- ²⁵ Pauli N, Barrios E, Conacher AJ, Oberthur T. 2012. Farmer knowledge of the relationships among soil macrofauna, soil quality, and tree species in a small holder agroforestry system of western Honduras. *Geoderma* 189–190:186–198.
- ²⁶ Tittonell P, Muriuki A, Shepherd KD, Mugendi D, Kaizzi KC, Okeyo J, Vanlauwe B. 2010. The diversity of rural livelihoods and their influence on soil fertility in agricultural systems of East Africa: a typology of smallholder farms. *Agricultural Systems* 103:83–97.
- ²⁷ Young A. 1997. Agroforestry for soil management. Edn. 2. Wallingford, UK: CAB International.
- ²⁸ Van Noordwijk M, Garrity DP. 1995. Nutrient use efficiency in agroforestry systems. In: *Potassium in Asia: balanced fertilization to increase and sustain agricultural production*. Basel, Switzerland: International Potash Institute. pp 245–279.

http://worldagroforestry.org/regions/southeast_asia/publications?do=view_pub_detail&pub_no=PP 0034-04

²⁹ Cannell MGR, van Noordwijk M, Ong CK. 1996. The central agroforestry hypothesis: the trees must acquire resources that the crop would not otherwise acquire. *Agroforestry Systems* 34:27–31.

³⁰ As note 8

- ³¹ Buresh RJ, Sánchez PA, Calhoun F, Hatfield J, Bigham JM. 1996. *Replenishing soil fertility in Africa*. Madison WI, USA: Soil Science Society of America.
- ³² Buresh RJ, Tian G. 1998. Soil improvement by trees in sub-Saharan Africa. In: Nair PKR, Latt CR, eds. Directions in tropical agroforestry research. Heidelberg, Germany: Springer Netherlands. pp 51–76.
- ³³ Shepherd KD, Ohlsson E, Okalebo JR, Ndufa JK. 1996. Potential impact of agroforestry on soil nutrient balances at the farm scale in the East African Highlands. *Fertilizer Research* 44:87–99.
- ³⁴ Coe R, Sinclair F, Barrios E. 2014. Scaling up agroforestry requires research 'in' rather than 'for' development. *Current Opinion in Environmental Sustainability* 6:73–77.
- ³⁵ Shepherd KD, Ndufa JK, Ohlsson E, Sjogren H, Swinkels R. 1997. Adoption potential of hedgerow intercropping in maize-based cropping systems in the highlands of western Kenya. I. Background and agronomic evaluation. *Experimental Agriculture* 33:197–223.
- ³⁶ Swinkels RA, Franzel SC, Shepherd KD, Ohlsson EL, Ndufa JK. 1997. The economics of short rotation improved fallows: evidence from areas of high population density in western Kenya. *Agricultural Systems* 55:99–121.
- ³⁷ Shepherd KD, Soule MJ. 1998. Economic and ecological impacts of soil management on west Kenyan farms: a dynamic simulation model. *Agriculture, Ecosystems and Environment* 71:131–146.
- ³⁸ Ndufa JK, Shepherd KD, Buresh RJ, Jama B. 1999. Nutrient uptake and growth of young trees in a Pdeficient soil: Tree species and phosphorus effects. *Forest Ecology and Management* 122:231–241.
- ³⁹ Sjögren H, Shepherd KD, Karlsson A. 2010. Effects of improved fallow with Sesbania sesban on maize productivity and Striga hermonthica infestation in Western Kenya. Journal of Forestry Research 21:379–386.
- ⁴⁰ Bayala J, Balesdent J, Marol C, Zapata F, Teklehaimanot Z, Ouedraogo SJ. 2007. Relative contribution of trees and crops to soil carbon content in a parkland system in Burkina Faso using variations in natural 13C abundance. In: Bationo A, Waswa B, Kihara J, Kimetu J, eds. Advances in integrated soil fertility management in sub-Saharan Africa: challenges and opportunities. Dordrecht, The Netherlands: Springer Netherlands. pp 161–169.
- ⁴¹ Van Noordwijk M, Cadisch G, Ong CK, eds. 2004. *Belowground interactions in tropical agroecosystems*. Wallingford UK: CAB International.
- ⁴² Barrios E, Buresh RJ, Sprent JI. 1996a. Organic matter in soil particle size and density fractions from maize and legume cropping systems. *Soil Biology & Biochemistry* 28:185–193.
- ⁴³ Meijboom FW, Hassink J, van Noordwijk M. 1995 Density fractionation of soil macro-organic matter using silica suspensions. Soil Biology and Biochemistry 27:1109–1111.
- 44 As note 43
- ⁴⁵ Barrios E, Buresh RJ, Sprent JI. 1996b. Nitrogen mineralization in density fractions of soil organic matter from maize and legume cropping systems. *Soil Biology and Biochemistry* 28:1459–1465.
- ⁴⁶ Barrios E, Kwesiga F, Buresh RJ, Sprent JI. 1997. Light fraction soil organic matter and available nitrogen following trees and maize. *Soil Science Society of America Journal* 61:826–831.
- ⁴⁷ Barrios E, Kwesiga F, Buresh RJ, Sprent JI, Coe R. 1998. Relating preseason soil nitrogen to maize yield in tree legume-maize rotations. *Soil Science Society of America Journal* 62:1604–1609.

48 As note 47

- ⁴⁹ Phiri S, Barrios E, Rao IM, Singh BR. 2001. Changes in soil organic matter and phosphorus fractions under planted fallows and a crop rotation system on a Colombian volcanic-ash soil. *Plant and Soil* 231:211–223.
- ⁵⁰ Kwesiga F, Franzel S, Place F, Phiri D, Simwanza CP. 1999. *Sesbania sesban* improved fallows in eastern Zambia: Their inception, development and farmer enthusiasm. *Agroforestry Systems* 47:49–66.
- ⁵¹ Barrios E, Cobo JG, Rao IM, Thomas RJ, Amezquita E, Jimenez JJ, Rondon MA. 2005. Fallow management for soil fertility recovery in tropical Andean agroecosystems in Colombia. *Agriculture, Ecosystems and Environment* 110:29–42.
- ⁵² Nair PR, Buresh RJ, Mugendi DN, Latt CR. 1999. Nutrient cycling in tropical agroforestry systems: myths and science. Agroforestry in sustainable agricultural systems. Boca Raton FL, USA: CRC Press; Totnes, UK: Lewis Publishers.

53 As note 42

- ⁵⁴ van Noordwijk M, Lawson G, Groot JJR, Hairiah K. 1996. Root distribution in relation to nutrients and competition. In: Ong CK, Huxley PA, eds. *Tree-crop interactions: a physiological approach*. Wallingford, UK: CAB International. pp 319–364.
- 55 As note 42
- ⁵⁶ van Noordwijk M. 1999. Nutrient cycling in ecosystems versus nutrient budgets of agricultural systems. Nutrient Cycles and Nutrient Budgets in Global Agro-ecosystems. Wallingford, UK: CAB International. pp 1–26.
- ⁵⁷ Stoorvogel JJ, Smaling EMA. 1990. *Assessment of soil nutrient depletion in Sub- Saharan Africa, 1983–2000.* Wageningen, The Netherlands: Winand Staring Centre for Integrated Soil and Water Research.
- ⁵⁸ Cobo JC, Dercon G, Cadisch G. 2010. Nutrient balances in African land use systems across different spatial scales: a review of approaches, challenges and progress. *Agriculture, Ecosystems and Environment* 136:1–15.
- ⁵⁹ Hairiah K, van Noordwijk M, Cadisch G. 2000. Crop yield, C and N balance of three types of cropping systems on an Ultisol in Northern Lampung. NJAS-Wageningen Journal of Life Sciences 48:3–17.
- ⁶⁰ Mafongoya PL, Kuntashula E, Sileshi G. 2006. *Managing soil fertility and nutrient cycles through fertilizer trees in southern Africa*. Biological Approaches to Sustainable Soil Systems. Abingdon, UK: Taylor & Francis. pp 273–289.
- ⁶¹ van Noordwijk M, Cadisch G. 2002. Access and excess problems in plant nutrition. *Plant and Soil* 247:25– 39.
- 62 As note 5
- 63 As note 38
- ⁶⁴ Jama B, Buresh RJ, Ndufa JK, Shepherd KD. 1998. Vertical distribution of roots and soil nitrate: tree species and phosphorus effects. *Soil Science Society of America Journal* 62:280–286.
- ⁶⁵ Shepherd G, Buresh RJ, Gregory PJ. 2000. Land use effects the distribution of soil inorganic nitrogen in smallholder production systems in Kenya. *Biology and Fertility of Soils* 31:348–355.
- ⁶⁶ Shepherd G, Buresh RJ, Gregory PJ. 2001. Inorganic soil nitrogen distribution in relation to soil properties in smallholder maize fields in the Kenyan highlands. *Geoderma* 101:87–103.
- ⁶⁷ Hoang Fagerström MH, Nilsson SI, van Noordwijk M, Thai Phien, Olsson M, Hansson A, Svensson C. 2002. Does *Tephrosia candida* as fallow species, hedgerow or mulch improve nutrient cycling and prevent nutrient losses by erosion on slopes in northern Vietnam? *Agriculture, Ecosystems & Environment* 90:291–304.
- ⁶⁸ Akinnifesi FK, Makumba W, Sileshi G, Ajayi OC, Mweta D. 2007. Synergistic effect of inorganic N and P fertilizers and organic inputs from *Gliricidia sepium* on productivity of intercropped maize in Southern Malawi. *Plant and Soil* 294:203–217.
- ⁶⁹ UNEP 2012. Land health surveillance: an evidence-based approach to land ecosystem management. Illustrated with a case study in the West Africa Sahel. Nairobi, Kenya: United Nations Environment Programme. <u>http://www.unep.org/dewa/Portals/67/pdf/LHS_Report_lowres.pdf.</u>
- ⁷⁰ Hoang MH, Joshi L, van Noordwijk M. 2013. Participatory landscape appraisal (PaLA). In: van Noordwijk M, Lusiana B, Leimona B, Dewi S, Wulandari D, eds. *Negotiation-support toolkit for learning landscapes*. Bogor, Indonesia. World Agroforestry Centre (ICRAF) Southeast Asia Regional Program. pp 16–21.

- ⁷¹ Awiti AO, Walsh MG, Shepherd KD, Kinyamario J. 2008. Soil condition classification using infrared spectroscopy: a proposition for assessment of soil condition along a tropical forest-cropland chronosequence. *Geoderma* 143:73–84.
- ⁷² Vågen TG, Walsh MG, Shepherd KD. 2008. *Potential for carbon storage through rehabilitation of degraded lands in the Lake Baringo Basin*. Nairobi, Kenya: United Nations Environment Programme.
- ⁷³ ICRAF. 2006. Improved land management in the Lake Victoria basin: final report on the TransVic project. Occasional Paper 07. Nairobi, Kenya: World Agroforestry Centre (ICRAF).
- ⁷⁴ Rao MR, Nair PKR, Ong CK. 1998. Biophysical interactions in tropical agroforestry systems. In: Nair PKR, Latt CR, eds. Directions in tropical agroforestry research. Heidelberg, Germany: Springer Netherlands. pp 3–50.
- ⁷⁵ Ong CK, Black C, Wilson J, eds. 2015. Tree-crop interactions: agroforestry in a changing climate. Wallingford, UK: CAB International.
- ⁷⁶ Lott JE, Howard SB, Black CR, Ong CK. 2000a. Long term productivity of a *Grevillea robusta*-based agroforestry system in Kenya. I. Tree growth. *Forest Ecology and Management* 139:175–186.
- ⁷⁷ Lott JE, Howard SB, Black CR, Ong CK. 2000b. Long term productivity of a *Grevillea robusta*-based agroforestry system in Kenya II. Crop growth and system productivity. *Forest Ecology and Management* 139:187–201.
- ⁷⁸ Broadhead JS, Ong CK, Black CR. 2003. Tree phenology and water availability in semi-arid agroforestry systems. *Forest Ecology and Management* 180(1–3):61–73.
- ⁷⁹ Broadhead JS, Black CR, Ong CK. 2003. Tree leafing phenology and crop productivity in semi-arid agroforestry systems in Kenya. *Agroforestry Systems* 58(2):137–148.
- ⁸⁰ Muthuri CW, Ong CK, Mati BM, van Noordwijk M. 2005. Modelling the effects of leafing phenology on growth and water use by selected agroforestry tree species in semi-arid Kenya. *Land and Water Resources Research* 4:1–11.
- ⁸¹ Ketterings QM, van Noordwijk M, Bigham JM. 2002. Soil phosphorus availability after slash-and-burn fires of different intensities in rubber agroforests in Sumatra, Indonesia. *Agriculture, Ecosystems and Environment* 92:37–48.
- ⁸² Rodenburg J, Stein A, van Noordwijk M, Ketterings QM. 2003. Spatial variability of soil pH and phosphorus in relation to soil run-off following slash-and-burn land clearing in Sumatra, Indonesia. *Soil and Tillage Research* 71:1–14.
- ⁸³ van Noordwijk M, Mulyoutami E, Sakuntaladewi N, Agus F. 2008. Swiddens in transition: shifted perceptions on shifting cultivators in Indonesia. Bogor, Indonesia: World Agroforestry Centre (ICRAF) Southeast Asia Regional Program.
- ⁸⁴ Bayala J, van Noordwijk M, Lusiana B, Kasanah N, Teklehaimanot Z, Ouedraogo SJ. 2008. Separating the tree-soil-crop interactions in agroforestry parkland systems in Saponé (Burkina Faso) using WaNuLCAS. Advances in Agroforestry 4:296–308.
- ⁸⁵ Bayala J, Bazié HR, Sanou J. 2013. Competition and facilitation-related factors impacts on crop performance in an agro-forestry parkland system in Burkina Faso. *African Journal of Agricultural Research* 8:5303–5310.
- ⁸⁶ Coulibaly YN, Mulia R, Sanou J, Zombre N, Bayala J, Kalinganire A, van Noordwijk M. 2014. Crop production under different rainfall and management conditions in agroforestry parkland systems in Burkina Faso: observations and simulation with WaNuLCAS model. *Agroforestry Systems* 88:13–28.
- 87 As note 83
- ⁸⁸ Agus F, Farida, van Noordwijk M, eds. 2004. Hydrological impacts of forest, agroforestry and upland cropping as a basis for rewarding environmental service providers in Indonesia. Proceedings of a workshop in Padang and Singkarak, West Sumatra, Indonesia, 25-28 February 2004. Bogor, Indonesia: World Agroforestry Centre (ICRAF) Southeast Asia Regional Program.
- ⁸⁹ Fonte SJ, Barrios E, Six J. 2010. Earthworms, soil fertility and aggregate-associated soil organic matter dynamics in the Quesungual agroforestry system. *Geoderma* 155:320–328.

- ⁹⁰ Malmer A, van Noordwijk M, Bruijnzeel LA, Bonell M. 2009. Effects of shifting cultivation and forest fire. In: Bonell M, Brijinzeel LA, eds. Forest-water-people in the humid tropics: past, present and future hydrological research for integrated land and water management. International Hydrology Series. Joint UNESCO International Hydrological Programme and International Union of Forestry Research Organizations symposium and workshop, Universiti Kebangsaan Malaysia, 30 July–4 August 2000. Cambridge, UK: Cambridge University Press. pp 533–560.
- 91 As note 87
- ⁹² Ong C, Black CR, Wilson J, Muthuri C, Bayala J, Jackson NA. 2014. Agroforestry: hydrological impacts. In: van Alfen NK, ed. *Encyclopedia of agriculture and food systems*. Amsterdam, The Netherlands: Elsevier Science.
- ⁹³ Radersma S, Ong CK. 2004. Spatial distribution of root length density and soil water of linear agroforestry systems in sub-humid Kenya: implications for agroforestry models. *Forest ecology and management* 188:77–89.
- ⁹⁴ van Noordwijk M, Martikainen P, Bottner P, Cuevas E, Rouland C, Dhillion SS. 1998. Global change and root function. *Global Change Biology* 4:759–772.
- ⁹⁵ Jonsson K. 1995. Agroforestry in dry savanna areas in Africa: interactions between trees, soils and crops. Thesis. Umea, Sweden: Swedish University of Agricultural Sciences.
- ⁹⁶ Bayala J, Balesdent J, Marol C, Zapata F, Teklehaimanot Z, Ouedraogo SJ. 2006. Relative contribution of trees and crops to soil carbon content in a parkland system in Burkina Faso using variations in natural 13C abundance. *Journal Nutrient Cycling in Agroecosystems* 76:193–201.
- ⁹⁷ Bayala J, Sanou J, Teklehaimanot Z, Kalinganire A, Ouédraogo SJ. 2014. Parklands for buffering climate risk and sustaining agricultural production in the Sahel of West Africa. *Current Opinion in Environmental Sustainability* 6:28–34.
- ⁹⁸ Rowe EC, Hairiah K, Giller KE, Van Noordwijk M, Cadisch G. 1999. Testing the safety-net role of hedgerow tree roots by 15 N placement at different soil depths. In: Auclair D, Dupraz C, eds. Agroforestry for sustainable land-use fundamental research and modelling with emphasis on temperate and Mediterranean applications. Dordrecht, The Netherlands: Springer. pp 81–93.
- ⁹⁹ Buresh RJ, Rowe EC, Livesley SJ, Cadisch G, Mafongoya P, van Noordwijk M, Ong CK. 2004. Opportunities for capture of deep soil nutrients. In: van Noordwijk M, Cadish G, Ong CK, eds. In: van Noordwijk M, Cadisch G, Ong CK, eds. *Belowground interactions in tropical agroecosystems: concepts and models with multiple plant components*. Wallingford, UK: CAB International. pp 109–125.
- ¹⁰⁰ Rowe EC, van Noordwijk M, Suprayogo D, Hairiah K, Giller KE, Cadisch G. 2001. Root distributions partially explain ¹⁵N uptake patterns in *Gliricidia* and *Peltophorum* hedgerow intercropping systems. *Plant and Soil* 235:167–179.
- ¹⁰¹ Burgess SS, Adams MA, Turner NC, Ong CK. 1998. The redistribution of soil water by tree root systems. *Oecologia* 115:306–311.
- ¹⁰² Bayala J, Heng LK, van Noordwijk M, Ouedraogo SJ. 2008a. Hydraulic Lift study in two native tree species of agroforestry parklands of West African dry savanna. *Acta Oecologica* 34:370–378.
- ¹⁰³ Sileshi G, Akinnifesi FK, Debusho LK, Beedy T, Ajayi OC, Mong'omba S. 2010. Variation in maize yield gaps with plant nutrient inputs, soil type and climate across sub-Saharan Africa. *Field Crops Res*earch 11:1–13.
- ¹⁰⁴ Bayala J, Sileshi GW, Coe R, Kalinganire A, Tchoundjeu Z, Sinclair F, Garrity D. 2012. Cereal yield response to conservation agriculture practices in drylands of West Africa: a quantitative synthesis. *Journal of Arid Environments* 78:13–25.
- ¹⁰⁵ Sileshi GW, Arshad MA, Konaté S, Nkunika PO. 2010. Termite-induced heterogeneity in African savanna vegetation: mechanisms and patterns. *Journal of Vegetation Science* 21:923–937.
- ¹⁰⁶ van Noordwijk M, Lusiana B. 1999. WaNuLCAS, a model of water, nutrient and light capture in agroforestry systems. *Agroforestry Systems* 43:217–242.
- ¹⁰⁷ Radersma S, Lusiana B, van Noordwijk M. 2005. Simulation of soil drying induced phosphorus deficiency and phosphorus mobilization as determinants of maize growth near tree lines on a Ferralsol. *Field Crops Research* 91:171–184.

¹⁰⁸ As note 85

¹⁰⁹ As note 87

¹¹⁰ Matthews R, van Noordwijk M, Gijsman AJ, Cadisch G. 2004. Models of belowground interactions: their validity, applicability and beneficiaries. In: van Noordwijk M, Cadisch G, Ong CK, eds. *Belowground interactions in tropical agroecosystems: concepts and models with multiple plant components.* Wallingford, UK: CAB International. pp 41–60.

¹¹¹ Boffa JM, Taonda SB, Dickey JB & Knudson DM. 2000. Field-scale influence of karité (*Vitellaria paradoxa*) on sorghum production in the Sudan zone of Burkina Faso. *Agroforestry Systems* 49(2):153–175.

¹¹² As note 55

- ¹¹³ Van Roode M. 2000. The effects of vegetative barrier strips on surface runoff and soil erosion in Machakos, Kenya. A statistical versus a spatial modelling approach. Thesis. Utrecht, The Netherlands: University of Utrecht.
- ¹¹⁴ Cohen MJ, Shepherd KD, Walsh MG. 2005. Empirical reformulation of the universal soil loss equation for erosion risk assessment in a tropical watershed. *Geoderma* 124:235–252.
- ¹¹⁵ Verbist B, Poesen J, van Noordwijk M, Widianto, Suprayogo D, Agus F, Deckers J. 2010. Factors affecting soil loss at plot scale and sediment yield at catchment scale in a tropical volcanic agroforestry landscape. *Catena* 80:34–46.
- ¹¹⁶ van Noordwijk M, van Roode M, McCallie EL, Lusiana B. 1998c. Erosion and sedimentation as multiscale, fractal processes: implications for models, experiments and the real world. In: Vries FP, Agus F, Kerr J, eds. *Soil erosion at multiple scales: principles and methods for assessing causes and impacts.* Wallingford, UK: CAB International. pp 223–253.
- ¹¹⁷ Ranieri SBL, Stirzaker R, Suprayogo D, Purwanto E, de Willigen P, van Noordwijk M. 2004. Managing movements of water, solutes and soil: from plot to landscape scale. In: van Noordwijk M, Cadisch G, Ong CK, eds. *Belowground interactions in tropical agroecosystems: concepts and models with multiple plant components*. Wallingford, UK: CAB International. pp 329–347.
- ¹¹⁸ Swallow BM, Garrity DP, van Noordwijk M. 2002. The effects of scales, flows and filters on property rights and collective action in watershed management. *Water Policy* 3:457–474.
- ¹¹⁹ Cohen MJ, Brown MT, Shepherd KD. 2006. Estimating the environmental costs of soil erosion at multiple scales in Kenya using emergy synthesis. Agriculture. *Ecosystems and Environment* 114:249–269.
- ¹²⁰ van Noordwijk M, Cadisch G, Ong CK. 2004. Challenges for the next decade of research on belowground interactions in tropical agroecosystems: client-driven solutions at landscape scale. In: van Noordwijk M, Cadisch G, Ong CK, eds. *Belowground interactions in tropical agroecosystems: concepts and models with multiple plant components.* Wallingford, UK: CAB International. pp 365–379.
- ¹²¹ van Noordwijk M, Widianto, Heinen M, Hairiah K. 1991. Old tree root channels in acid soils in the humid tropics: important for crop root penetration, water infiltration and nitrogen management. *Plant Soil* 134:37–44.
- ¹²² Hairiah K, Sulistyani H, Suprayogo D, Purnomosidhi P, Widodo RH, van Noordwijk M. 2006. Litter layer residence time in forest and coffee agroforestry systems in Sumberjaya, West Lampung. *Forest Ecology and Management* 224:45–57.
- ¹²³ As note 90
- ¹²⁴ Kuyper TW, Cardoso IM, Onguene NA, Murniati VNM. 2004. Managing mycorrhiza in tropical multispecies agroecosystems. In: van Noordwijk M, Cadisch G, Ong CK, eds. *Belowground interactions in tropical agroecosystems: concepts and models with multiple plant components*. Wallingford, UK: CAB International. pp 243–261.
- ¹²⁵ Velasquez E, Pelosi C, Brunet D, Grimaldi M, Martins M, Rendeiro AC, Barrios E, Lavelle P. 2007. This ped is my ped: Visual separation and near infrared spectra allow determination of the origins of soil macroaggregates. *Pedobiologia* 51:75–87.
- ¹²⁶ Suprayogo D, van Noordwijk M, Hairiah K, Cadisch G. 2002. The inherent 'safety-net' of an Acrisol: measuring and modelling retarded leaching of mineral nitrogen. *European Journal of Soil Science* 53:185–194.
- ¹²⁷ Van Noordwijk M, Lusiana B, Khasanah N, Mulia R. 2011. WaNuLCAS version 4.0, Background on a Model of Water Nutrient and Light Capture in Agroforestry Systems. Bogor, Indonesia: World Agroforestry Centre (ICRAF).
- ¹²⁸ Ilstedt U, Tobella AB, Bazié HR, Bayala J, Verbeeten E, Nyberg G, Sanou J, Benegas L, Murdiyarso D, Laudon H, Sheil D. 2016. Intermediate tree cover can maximize groundwater recharge in the seasonally dry tropics. *Scientific Reports* 6. p 21930.

- ¹²⁹ Susilo FX, Neutel AM, van Noordwijk M, Hairiah K, Brown G, Swift MJ, Ong CK. 2004. Soil biodiversity and food webs. In: van Noordwijk M, Cadisch G, Ong CK, eds. *Belowground interactions in tropical agroecosystems: concepts and models with multiple plant components*. Wallingford, UK: CAB International. pp 285–307.
- ¹³⁰ Barrios E. 2007. Soil biota, ecosystem services and land productivity. *Ecological Economics* 64:269–285.
 ¹³¹ As note 25
- ¹³² Wu T, Ayres E, Bardgett RD, Wall DH, Garey JR. 2011 Molecular study of worldwide distribution and diversity of soil animals. *Proceedings of the National Academies of Science* 108(43):17720–17725.
- ¹³³ As note 125
- ¹³⁴ As note 90
- ¹³⁵ Barrios E, Sileshi GW, Shepherd K, Sinclair F. 2012b. Agroforestry and soil health: linking trees, soil biota and ecosystem services. In: Wall DH, ed. Oxford Handbook of Soil Ecology and Ecosystem Services. Oxford, UK: Oxford University Press. pp 315–330.
- ¹³⁶ Shepherd K, Walsh M. 2006, July. Infrared spectroscopy: new technology for boosting agricultural productivity and monitoring environment in developing countries. Proceedings of the 18th World Congress of Soil Science, 9–15 July 2006, Philadelphia, USA. <u>www.ldd.go.th/18wcss/techprogram/P15007.htm</u>.
- ¹³⁷ Tata HL, van Noordwijk M, Summerbell R, Werger MJA. 2010. Limited response to nursery-stage mycorrhiza inoculation of Shorea seedlings planted in rubber agroforest in Jambi, Indonesia. *New Forests* 39:51–74.
- ¹³⁸ Maestre FT, Quero JL, Gotelli NJ, Escudero A, Ochoa V, Delgado-Baquerizo M, García-Gómez M, Bowker MA, Soliveres S, Escolar C, García-Palacios P, Berdugo M, Valencia E, Gozalo B, Gallardo A, Aguilera L, Arrendondo T, Blones J, Boeken B, Bran D, Conceição AA, Cabrera O, Chaieb M, Derak M, Eldridge DF, Espinosa CI, Florentino A, Gaitán J, Gabriel Gatica M, Ghiloufi W, Gómez-González S, Gutiérrez JR, Hernández RM, Huang X, Huber-Sannwald E, Jankju M, Miriti M, Monerris J, Mau RL, Morici E, Naseri K, Ospina A, Polo V, Prina A, Pucheta E, Ramíerez-Collantes DA, Romão R, Tighe M, Torres-Díaz C, Val J, Veiga JP, Wang D, Zaady E. 2012. Plant species richness and ecosystem multifunctionality in global drylands. *Science* 335(6065):214–218.
- ¹³⁹ Pauli N, Barrios E, Conacher AJ, Oberthur T. 2011. Soil macrofauna in agricultural landscapes dominated by the Quesungual slash-and-mulch agroforestry system, western Honduras. *Applied Soil Ecology* 47:119–132.
- 140 As note 136
- ¹⁴¹ Pauli N, Oberthur T, Barrios E, Conacher A. 2010. Fine-scale spatial and temporal variation in earthworm surface casting activity in agroforestry fields, western Honduras. *Pedobiologia* 53:127–139.
- ¹⁴² Ordonez JC, Luedeling E, Kindt R, Tata HL, Harja D, Jamnadass R, van Noordwijk M. 2014. Constraints and opportunities for tree diversity management along the forest transition curve to achieve multifunctional agriculture. *Current Opinion in Environmental Sustainability* 5:54–60.
- 143 As note 136
- ¹⁴⁴ Swift MJ, Izac AM, van Noordwijk M. 2004. Biodiversity and ecosystems services in agricultural landscapes: are we asking the right questions? *Agriculture, Ecosystems and Environment* 104:113– 134.
- ¹⁴⁵ Giller KE, Bignell DE, Lavelle P, Swift MJ, Barrios E, Moreira F, van Noordwijk M, Barois I, Karanja N, Huising J. 2005. Soil biodiversity in rapidly changing tropical landscapes: scaling down and scaling up. In: Bardgett RD, Usher MB, Hopkins DW. *Biological diversity and function in soils*. Cambridge, UK: Cambridge University Press. pp 295–318.
- ¹⁴⁶ Sileshi GW, Kuntashula E, Matakala P, Nkunika PO. 2008. Farmers' perceptions of tree mortality, pests and pest management practices in agroforestry in Malawi, Mozambique and Zambia. *Agroforestry Systems* 72:87–101.
- 147 As note 26
- ¹⁴⁸ Palm CA, van Noordwijk M, Woomer PL, Alegre JC, Arevalo L, Castilla CE, Sitompol SM. 2005. Carbon losses and sequestration after land use change in the humid tropics. In: Vosti SA, Sanchez PA, eds. *Slash-and-burn agriculture: the search for alternatives*. New York, USA: Columbia University Press.
- ¹⁴⁹ Van Noordwijk M, Cerri C, Woomer PL, Nugroho K, Bernoux M. 1997. Soil carbon dynamics in the humid tropical forest zone. *Geoderma* 79:187–225.

- ¹⁵⁰ Sitompul SM, Hairiah K, Cadisch G, van Noordwijk M. 2000. Dynamics of density fractions of macroorganic matter after forest conversion to sugarcane and woodlots, accounted for in a modified Century model. NJAS-Wageningen Journal of Life Sciences 48:61–73.
- ¹⁵¹ Albrecht A, Kandji ST. 2003. Carbon sequestration in tropical agroforestry systems. *Agriculture, Ecosystems and Environment* 99:15–27.
- ¹⁵² Van Noordwijk M, Rahayu S, Hairiah K, Wulan YC, Farida A, Verbist B. 2002. Carbon stock assessment for a forest-to-coffee conversion landscape in Sumber-Jaya (Lampung, Indonesia): from allometric equations to land use change analysis. *Science in China series C Life Sciences* 45 Supp. pp 75–86.
- ¹⁵³ Van Noordwijk M. 2014. Avoided land degradation and enhanced soil C storage: is there a role for carbon markets? In: Banwart SA, Noellemeyer E, Milne E, eds. *Soil carbon science, management and policy for multiple benefits*. SCOPE Series Vol 71. Wallingford, UK: CAB International.
- ¹⁵⁴ Van Noordwijk M, Goverse T, Ballabio C, Banwart S, Bhattacharyya T, Goldhaber M, Nikolaidis N, Noellemeyer E, Zhao Y. 2014a. Soil organic carbon transition curves: reversal of land degradation through management of soil organic matter for multiple benefits. In: Banwart SA, Noellemeyer E, Milne E, eds. Soil carbon science, management and policy for multiple benefits. SCOPE Series Vol 71. Wallingford, UK: CAB International.
- ¹⁵⁵ Verchot LV, van Noordwijk M, Kandji S, Tomich T, Ong C, Albrecht A, Palm C. 2007. Climate change: linking adaptation and mitigation through agroforestry. *Mitigation and Adaptation Strategies for Global Change* 12:901–918.
- ¹⁵⁶ Verchot LV, Hutabarat L, Hairiah K, van Noordwijk M. 2006. Nitrogen availability and soil N₂O emissions following conversion of forests to coffee in southern Sumatra. *Global Biogeochemistry Cycles* 20:GB4008. DOI: 10.1029/2005GB002469.
- ¹⁵⁷ Rosenstock TS, Mpanda M, Rioux J, Aynekulua E, Kimaro AA, Neufeldt H, Shepherd KD, Luedeling E. 2014. Targeting conservation agriculture in the context of livelihoods and landscapes. *Agriculture Ecosystems & Environment* 187:47–51.
- ¹⁵⁸ Shepherd KD, Walsh MG. 2000. Sensing soil quality: the evidence from Africa. Natural Resource Problems, Priorities and Policies Programme Working Paper 2000–1. Nairobi, Kenya: International Centre for Research in Agroforestry.
- 159 As note 158
- ¹⁶⁰ Vågen TG, Shepherd KD, Walsh MG. 2006. Sensing landscape level change in soil fertility following deforestation and conversion in the highlands of Madagascar using Vis-NIR spectroscopy. *Geoderma* 133:281–294.
- ¹⁶¹ As note 72
- ¹⁶² Shepherd KD, Walsh MG. 2002. Development of reflectance spectral libraries for characterization of soil properties. *Soil Science Society of America Journal* 66:988–998.
- ¹⁶³ Brown DJ, Shepherd KD, Walsh MG, Dewayne Mays M, Reinsch TG. 2006. Global soil characterization with VNIR diffuse reflectance spectroscopy. *Geoderma* 132:273–290.
- ¹⁶⁴ Terhoeven-Urselmans T, Vågen TG, Spaargaren O, Shepherd KD. 2010. Prediction of soil fertility properties from a globally distributed soil mid-infrared spectral library. *Soil Science Society of America Journal* 74:1792–1799.
- ¹⁶⁵ Shepherd KD, Walsh MG. 2007. Infrared spectroscopy: enabling an evidence-based diagnostic surveillance approach to agricultural and environmental management in developing countries. *Journal of Near Infrared Spectroscopy* 15:1–19.
- ¹⁶⁶ Nocita M, Stevens A, van Wesemael B, Brown DJ, Shepherd KD, Towett E, Vargase R, Montanarella L. 2014. Soil spectroscopy: an opportunity to be seized. *Global Change Biology*. DOI:10.1111/gcb.12632.
- ¹⁶⁷ Shepherd KD. 2007. Saving Africa's soils: science and technology for improved soil management in Africa. Nairobi, Kenya: World Agroforestry Centre (ICRAF).
- ¹⁶⁸ As note 161
- ¹⁶⁹ As note 72
- ¹⁷⁰ As note 70
- ¹⁷¹ Muhati SI, Shepherd KD, Gachene CK, Mburu MW, Jones R, Kironchi GO, Sila A. 2011. Diagnosis of soil nutrient constraints in small-scale groundnut (Arachis hyopaea L.) production systems of western Kenya using infrared spectroscopy. *Journal of Agricultural Science and Technology* A:111–127.

¹⁷² As note 114

- ¹⁷³ DeGraffenried JB, Shepherd KD, 2009. Rapid erosion modeling in a Western Kenya watershed using visible near infrared reflectance, classification tree analysis and 137Cesium. *Geoderma* 154(1–2):93– 100.
- ¹⁷⁴ Walsh M, Shepherd KD, Awiti A, Vågen TG. 2006. Land degradation surveillance: a spatial framework for characterization, research and development. Proceedings of the 18th World Congress of Soil Science, 9–15 July 2006, Philadelphia, USA. <u>www.ldd.go.th/18wcss/techprogram/P15007.htm</u>.
- ¹⁷⁵ Janik LJ, Skjemstad JO, Shepherd KD, Spouncer LR. 2007. The prediction of soil carbon fractions using mid-infrared-partial least square analysis. *Journal of Australian Soil Research* 45:73–81.
- ¹⁷⁶ Mutuo PK, Shepherd KD, Albrecht A, Cadisch G. 2006. Prediction of carbon mineralization rates from different soil physical fractions using diffuse reflectance spectroscopy. *Soil Biology and Biochemistry* 38:1658–1664.
- ¹⁷⁷ Verchot LV, Dutaur L, Shepherd KD, Albrecht A. 2011. Organic matter stabilization in soil aggregates: understanding the biogeochemical mechanisms that determine the fate of carbon inputs in soils. *Geoderma* 161:182–193.
- ¹⁷⁸ Kamau-Rewe M, Rasche F, Cobo JG, Dercon G, Shepherd KD, Cadisch G. 2011. Generic prediction of soil organic carbon in Alfisols using diffuse reflectance Fourier transformed mid-infrared spectroscopy. *Soil Science Society of America Journal* 75:2358–2360.
- ¹⁷⁹ Shepherd KD, Palm CA, Gachengo CN, Vanlauwe B. 2003. Rapid characterization of organic resource quality for soil and livestock management in tropical agroecosystems using near infrared spectroscopy. *Agronomy Journal* 95:1314–1322.
- ¹⁸⁰ Vanlauwe B, Gachengo C, Shepherd K, Barrios E, Cadisch G, Palm CA. 2005. Laboratory validation of a resource quality-based conceptual framework for organic matter management. *Soil Science Society* of America Journal 69:1135–1145.
- ¹⁸¹ Shepherd KD, Vanlauwe B, Gachengo CN, Palm CA. 2005. Decomposition and mineralization of organic residues predicted using near infrared spectroscopy. *Plant and Soil* 277(1–2):315–333.
- ¹⁸² Tscherning K, Barrios E, Lascano CE, Peters M, Schultze-Kraft R. 2005. Effects of sample post-harvest treatment on aerobic decomposition and anaerobic in-vitro digestion of tropical legumes with contrasting quality. *Plant and Soil* 269:159–170.
- ¹⁸³ Tscherning K, Lascano CE, Barrios, E., Schultze-Kraft, R., Peters, M. 2006. The effect of mixing prunings of two tropical shrub legumes (Calliandra houstoniana and Indigofera zollingeriana) with contrasting quality on N release in the soil and apparent N degradation in the rumen. *Plant and Soil* 280:357– 368.
- ¹⁸⁴ Towett EK, Alex M, Shepherd KD, Polreich S, Aynekulu E, Maass BL. 2013a. Applicability of near-infrared reflectance spectroscopy (NIRS) for determination of crude protein content in cowpea (Vigna unguiculata) leaves. *Food Science and Nutrition* 1:45–53.
- ¹⁸⁵ Shepherd KD, Walsh MG. 2007. Infrared spectroscopy: enabling an evidence-based diagnostic surveillance approach to agricultural and environmental management in developing countries. *Journal of Near Infrared Spectroscopy* 15(1):1–19.
- ¹⁸⁶ Shepherd KD, Vågen TG, Gumbricht T, Walsh MG. 2008. Land degradation surveillance: quantifying and monitoring land degradation. in sustainable land management sourcebook. Washington DC, USA: World Bank. pp 141–147.

- ¹⁸⁹ Vågen TG, Winowiecki LA., Tondoh JE. 2013. The land degradation surveillance framework field guide. Version 4. Nairobi, Kenya: World Agroforestry Centre (ICRAF).
- ¹⁹⁰ UNEP 2012. Sahel atlas of changing landscapes: tracing trends and variations in vegetation cover and soil condition. Nairobi, Kenya: United Nations Environment Programme. <u>http://www.unep.org/dewa/Portals/67/pdf/Sahel_Atlas_lowres.pdf</u>
- ¹⁹¹ Sanchez PA, Ahamed S, Carré F, Hartemink AE, Hempel J, Huising J, Lagacherie P, McBratney AB, McKenzie NJ, de Lourdes Mendonça-Santos M, Minasny B. 2009. Digital soil map of the world. *Science 325*(5941):680–681.
- ¹⁹² Vågen TG, Shepherd KD, Walsh MG, Winowiecki L, Desta LT, Tondoh JE. 2010. AfSIS technical specifications. Soil health surveillance. Version 1.0. Nairobi, Kenya: World Agroforestry Centre (ICRAF).

¹⁸⁷ As note 70

¹⁸⁸ As note 70

- ¹⁹³ Towett EK, Shepherd KD, Cadisch G. 2013b. Quantification of total element concentrations in soils using total X-ray fluorescence spectroscopy (TXRF). *Science of the Total Environment* 463–464:374–388.
- ¹⁹⁴ Vågen TG, Davey FA, Shepherd KD. 2012. Land health surveillance: mapping soil carbon in rangelands. In Nair PKR, Garrity D, eds. Agroforestry: the future of global land use. Dordrecht, The Netherlands: Springer. pp 455–462.
- ¹⁹⁵ Vågen TG, Winowiecki LA, Abegaz A, Hadgu KM. 2013. Landsat-based approaches for mapping of land degradation prevalence and soil functional properties in Ethiopia. *Remote Sensing of Environment* 134:266–275.
- ¹⁹⁶ Barrios E, Coutinho HL, Medeiros CA. 2012. InPaC-S: participatory knowledge integration on indicators of soil quality: methodological guide. Nairobi, Kenya: World Agroforestry Centre (ICRAF); Brasilia, Brazil: Embrapa; Cali, Colombia: Centro Internacional de Agricultura Tropical.
- ¹⁹⁷ Tittonell P, Vanlauwe B, Leffelaar PA, Shepherd KD, Giller KE. 2005. Exploring diversity in soil fertility management of smallholder farms in western Kenya II. Within-farm variability in resource allocation, nutrient flows and soil fertility status. *Agriculture, Ecosystems and Environment* 110:166– 184.
- ¹⁹⁸ Tittonell P, Shepherd KD, Vanlauwe B, Giller KE 2008. Unravelling the effects of soil and crop management on maize productivity in smallholder agricultural systems of western Kenya: an application of classification and regression tree analysis. *Agriculture, Ecosystems and Environment* 123:137–150.

¹⁹⁹ As note 27

- ²⁰⁰ Tittonell P, Muriuki A, Klapwijk CJ, Shepherd KD, Coe R, Vanlauwe B. 2013. Soil heterogeneity and soil fertility gradients in smallholder agricultural systems of the East African Highlands. *Soil Science Society of America Journal* 77:525–538.
- ²⁰¹ Herrick JE, Urama KC, Karl JW, Boos J, Johnson MV, Shepherd KD, Hempel J, Bestelmeyer BT, Davies J, Guerra JL, Kosnik C, Kimiti DW, Losinyen Ekai A, Muller K, Norfleet L, Ozor N, Reinsch T, Sarukhan J, West LT. 2013. The global Land-Potential Knowledge System (LandPKS): supporting evidencebased, site-specific land use and management through cloud computing, mobile applications, and crowdsourcing. *Journal of Soil and Water Conservation* 68:5A–12A.
- ²⁰² Rosenstock TS, Mpanda M, Rioux J, Aynekulua E, Kimaro AA, Neufeldt H, Shepherd KD, Luedeling E. 2014. Targeting conservation agriculture in the context of livelihoods and landscapes. *Agriculture, Ecosystems and Environment* 187:47–51.
- ²⁰³ Van Noordwijk M, Wadman W. 1992. Effects of spatial variability of nitrogen supply on environmentally acceptable nitrogen fertilizer application rates to arable crops. *Netherlands Journal of Agricultural Science* 40:51–72.
- ²⁰⁴ Sanchez PA, Jama BA. 2001. Soil fertility replenishment takes off in East and Southern Africa. In: Vanlauwe B, Diels J, Sanginga N, Merckx R, eds. *Integrated plant nutrient management in sub-Saharan Africa: from concept to practice*. Wallingford, UK: CABI Publishing; Ibadan, Nigeria: International Institute of Tropical Agriculture. pp 23–45.
- ²⁰⁵ Sanchez PA. 2002. Soil fertility and hunger in Africa. *Science* (Washington) 295:2019–2020.
- ²⁰⁶ Ajayi OC, Akinnifesi FK, Sileshi G, Chakeredza S. 2007. Adoption of renewable soil fertility replenishment technologies in the southern African region: Lessons learnt and the way forward. *Natural Resources Forum* 31(4):306–317.
- ²⁰⁷ Place F, Franzel S, DeWolf J, Rommelse R, Kwesiga F, Niang A, Jama B. 2002. Agroforestry for soil-fertility replenishment: evidence on adoption processes in Kenya and Zambia. In: Barrett C, Place F, Aboud A, eds. *Natural resources management in African agriculture: understanding and improving current practices*. Wallingford UK: CABI International. pp 155–180.
- ²⁰⁸ Van Noordwijk M, Dijksterhuis G, van Keulen H. 1994. Risk management in crop production and fertilizer use with uncertain rainfall: how many eggs in which baskets. *Netherlands Journal of Agricultural Science* 42:249–269.

- ²¹⁰ As note 181
- ²¹¹ As note 71
- ²¹² Van Noordwijk M, Namirembe S, Catacutan DC, Williamson D, Gebrekirstos A. 2014b. Pricing rainbow, green, blue and grey water: tree cover and geopolitics of climatic teleconnections. *Current Opinion in Environmental Sustainability* 6:41–47.

²⁰⁹ As note 35

²¹³ As note 71

- ²¹⁴ Mbow C, van Noordwijk M, Prabhu R, Simons AJ. 2014b. Knowledge gaps and research needs concerning agroforestry's contribution to sustainable development goals in Africa. *Current Opinion in Environmental Sustainability* 6:162–170.
- ²¹⁵ Mbow C, Neufeldt H, van Noordwijk M, Minang P, Kowero G, Luedeling E. 2014a. Agroforestry solutions to address climate change and food security challenges in Africa. *Current Opinion in Environmental Sustainability* 6:61–67.
- ²¹⁶ Bernard F, van Noordwijk M, Luedeling E, Villamor GB, Gudeta S, Namirembe S. 2014 Social actors and unsustainability of agriculture. *Current Opinion in Environmental Sustainability* 6:155–161.